

HEAVY METAL AND METALLOID CONCENTRATION DYNAMICS IN MINE AND SURFACE WATERS IN THE VICINITY OF THE CHIPROVTSI AND MARTINOVO MINES, NORTHWESTERN BULGARIA

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Abstract

Chiprovtsi silver-lead and Martinovo iron mines represent the biggest mining area in Northwestern Bulgaria, which was operated till 1999. Their long-lived operation leads to proved pollution of the environment in the vicinity of the mines, especially water and soil. Seasonal monitoring of heavy metal (Cu, Zn, Cd and Pb) and metalloid (As and Sb) concentrations in mine, surface (river) and drinking waters was carried out during May and August 2006 to determine the level of contamination of the Chiprovtska Ogosta river basin resulting from the long-lived mining activity and whether these abandoned mines continue to be potential source for water pollution. This study proves significant As concentrations in mine (up to 170 µg/l) and surface waters (between 50 and 621 µg/l). The presence of other heavy metals, such as Cu, Cd, Zn and Pb, and metalloid – Sb is also recorded. Among them, Pb was found in considerable concentrations – up to 1456 µg/l during May 2006 sampling exhibiting great concentration variability between dry and wet sampling seasons. Sb is also determined in mine waters (up to 25 µg/l), but not exists in surface and drinking waters. Drinking waters are proved to be free of heavy metals and metalloids.

Key words: Chiprovtska Ogosta river basin, water pollution, arsenic, antimony, lead.

Περίληψη

Τα ορυχεία αργύρου-μολύβδου στο Chiprovtsi και σιδήρου στο Martinovo που λειτουργούσαν έως το 1999, αποτελούν την κύρια περιοχή μεταλλευτικής δραστηριότητας στη βορειοδυτική Βουλγαρία. Η μακροχρόνια λειτουργία τους είχε ως αποτέλεσμα εμφανή περιβαλλοντική ρύπανση, κυρίως στα ύδατα και στο έδαφος της

ευρύτερης περιοχής. Για να καθοριστεί το επίπεδο μόλυνσης της ποτάμιας λεκάνης Chiprovtska Ogosta από τη μακροχρόνια μεταλλευτική δραστηριότητα, αλλά και να προσδιοριστεί εάν τα σημερινά εγκαταλειμμένα μεταλλεία συνεχίζουν να αποτελούν πηγή ρύπανσης των υδάτων, πραγματοποιήθηκε εποχικός έλεγχος των συγκεντρώσεων βαρέων μετάλλων (Cu, Zn, Cd και Pb) και μεταλλοειδών (As και Sb) σε επιφανειακά (ποτάμια), πόσιμα ύδατα, καθώς και στα ύδατα των μεταλλείων κατά τη διάρκεια του Μαΐου και του Αυγούστου του 2006. Από τη μελέτη πιστοποιούνται σημαντικές συγκεντρώσεις As στα ύδατα των μεταλλείων (έως 170 $\mu\text{g/l}$) και στα επιφανειακά ύδατα (μεταξύ 50 και 621 $\mu\text{g/l}$). Επίσης, καταγράφεται η παρουσία άλλων βαρέων μετάλλων, όπως είναι τα Cu, Cd, Zn και Pb, και μεταλλοειδών – Sb. Ανάμεσα σε αυτά, ο Pb παρουσίασε υψηλές συγκεντρώσεις –έως 1456 $\mu\text{g/l}$ κατά τη διάρκεια της δειγματοληψίας του Μαΐου του 2006 εμφανίζοντας μεγάλη διακύμανση των συγκεντρώσεων μεταξύ ξηρών και υγρών περιόδων δειγματοληψίας. Επίσης, το Sb στα ύδατα των μεταλλείων παρουσίασε αξιοσημείωτες τιμές συγκεντρώσεων (μέχρι 25 $\mu\text{g/l}$), οι οποίες δεν παρατηρήθηκαν στα επιφανειακά και πόσιμα ύδατα. Τα πόσιμα ύδατα αποδεικνύεται ότι δεν φέρουν βαρέα μέταλλα και μεταλλοειδή.

Λέξεις κλειδιά: Ποτάμια λεκάνη Ogosta Chiprovtska, ρύπανση υδάτων, αρσενικό, αντιμόνιο, μόλυβδος.

1. Introduction

Mine waste tailing materials and mine waters are rich source of potentially toxic metallic and metalloid elements to the environment. Surface waters, which drain the tailing materials and wind transported fine tailing particles, create constant fluxes of toxic elements, which in turn get into river systems, soil and vegetation. Once reach the river system, pollutants can be transported to dozens of kilometers along the river downstream causing riverside area contamination far from the mining district. Although their concentrations decrease, due to dilution by surface waters, or as a result of adsorption by stream sediments, for some mobile elements, such as arsenic, the concentrations remain much higher compared to the maximum permissible level for surface waters.

The monitoring of the seasonal dynamics and concentrations of the potential pollutants in surface waters, which drain the tailings, and mine and river waters will help in prediction and prevention of possible river (irrigation) and drinking water contamination in concentrations above the guideline values. In the vicinity of the Chiprovtsi and Martinovo mines, there is a typical case of mine waters discharging into stream waters, as well as natural and clarified waters from the tailings collectors, which flow directly into Chiprovtska Ogosta river system. All these significantly contribute in Chiprovtska Ogosta river and riverside area pollution.

The terms “heavy metal” and “metalloid” are used in the paper to designate the elements Cu, Zn, Pb and Cd, and As and Sb respectively, on the basis of their general physical and chemical properties, as it is accepted in chemistry.

2. Geology and mining history of the Martinovo iron and Chiprovtsi silver-lead deposits

The Martinovo iron and Chiprovtsi silver-lead deposits are hosted in the low-grade metamorphic rocks (marbles and schists) of the Diabase Phyllitoid Complex (Precambrian - Early Ordovician age) (Kalvacheva 1986, Carrigan *et al.* 2003). The mineralizations in both deposits comprise of Mo-W (molybdenite-scheelite), Fe (magnetite-pyrrhotite-arsenopyrite-löllingite-pyrite-siderite), Ag-Pb (galena-tetrahedrite-Ag-Sb sulphosalts) and fluorite-barite-calcite (with minor constituent of cinnabar) (Atanassov and Pavlov 1983). Chiprovtsi and Martinovo mines are known to be exploited from Roman times, through Middle Ages, but most intensively from 1951 to 1999, when both mines ceased their activities after producing 4.79 Mt at 1.84 % Pb and 14.0 Mt at 30.7 % Fe (Milev *et al.* 1996). Remediation activities have been carried out in the region since 2000 year,

which covered not only the three tailing impoundments, but the numerous mine waste dumps too. At present all mine waste dumps are partially or completely remediated, except those in Rupski Dol, Lukina Padina and Dolich mine sections.

3. Investigation objectives

The current investigation aims to study the seasonal dynamics and distribution of the main pollutants (As, Sb, Pb, Cu, Zn and Cd) in mine, stream and Chiprovtska Ogosta river waters, as well as their environmental state regarding remediation activities, which were carried out in the mining area since 2000 year. Seasonal alteration of main pollutants concentrations is substantial assessing the remediation effectiveness and environmental stability of the river basin.

4. Materials and methods

Sampling took place during both wet (high water) (May 2006) and dry seasons (low water) (August 2006) in all mine sections and at the three tailing impoundments (Mechi Dol, Chiprovtsi and Golyam Bukovets) in the vicinity of the Martinovo and Chiprovtsi mines. Samples were collected from mine waters discharging from the main adits in the Perchinki, Mali Dol, Zhazhkov Dol, Gnili Dol, Velin Dol, Lukina Padina and Dolich mine sections and from stream waters in the Yavorov Dol, Sinya Voda and Rupski Dol mine sections, as well as samples from particular points along the Golyam Bukovets river and Chiprovtska Ogosta river downstream. Two samples of drinking water were collected from public fountains in the village of Zhelezna (Fig. 1). Sample locations were recorded with a handheld Garmin-GPS device. Physico-chemical parameters of waters (pH, T, TDS and EC) were measured in situ with Combo HANNA portable equipment (pH-meter and conductivity meter). All water samples were collected according to the international water sampling guidelines (WHO, 1996), ISO 5667-3:1994 and ISO 5667-14:1998. Each 100 ml water sample was collected in polyethylene bottle with 50 ml syringe; in situ filtrated through 0.45 μm pore diameter membrane filters, and acidified separately with supra pure acids: 5 ml HCl (for As and Sb), and 1 ml HNO_3 (for Pb, Cu, Zn and Cd). The concentrations of As, Sb, Pb, Cu, Zn and Cd in the samples were determined by inductively coupled plasma atomic emission spectroscopy using HR-ICP-AES Jobin Yvon Ultima 2 at the Geological Institute, BAS and ICP-AES equipment at the geochemistry laboratory, University of Mining and Geology, Sofia in agreement with ISO 11885:1996(E).

5. Results and discussion

5.1. Seasonal concentration variations of heavy metals and metalloids in mine waters

The measured in situ physico-chemical parameters of mine waters slightly differ during May 2006 and August 2006, but distinct increase of the measured temperature, total dissolved solids (TDS) and conductivity is observed during dry season for all sampled waters. Seasonal monitoring did not ascertain regular seasonal correlation of pH values for all mine waters. pH values range from 7.40 (Mali Dol, May 2006) up to 8.76 (Dolich, August 2006), but remain almost equal to neutral-near neutral pH, due to the carbonate-buffering by the host rocks (siderite marbles). This also prevents acid mine drainage (AMD) generation discharge from the mines and retains the sulphides dissolution ability low. The estimated total concentrations of heavy metals and metalloids in the waters collected in May 2006 are significantly higher than those collected in August 2006, except for the arsenic contents, which have reverse correlation and are relatively higher during August 2006 sampling (Table 1).

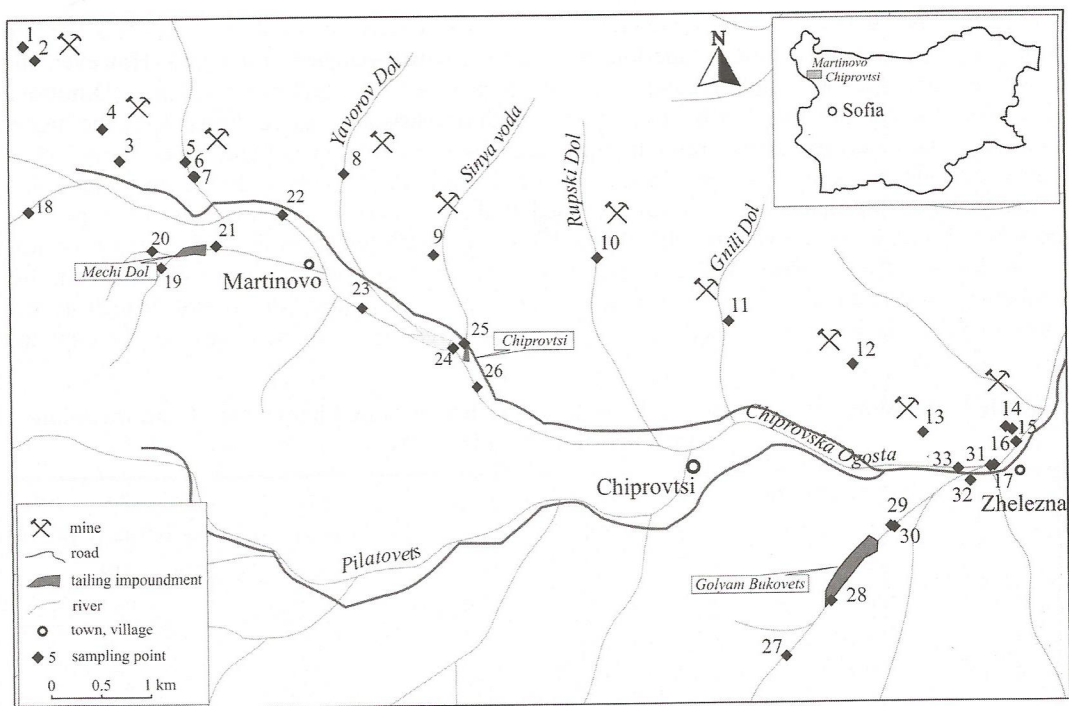


Figure 1 - Location of the Chiprovtska Ogosta river basin, showing the mining area, tailing impoundments and sampling points

The observed concentrations of each element show apparent correlation with the particular mineralization in each mine section. The highest arsenic concentrations are determined in the mine waters from Perchinki, Mali Dol and Zhazhkov Dol mine sections, where the ore is strongly enriched with arsenopyrite and löllingite. Similarly to this, the highest lead and antimony concentrations determined in waters from Sinya Voda, Gnili Dol, Lukina Padina and Dolich mine sections are related to the galena-tetrahedrite containing ore in Chiprovtsi mine.

Previous mine water investigations carried out by Kotsev (2001, 2003), when compared to this study, show considerable decrease in the lead, zinc, copper, cadmium and arsenic concentrations in mine waters at present, probably resulting from the cessation of mining activity.

However, the lack of data about antimony concentration in mine waters, especially from Lukina Padina and Dolich mine sections, did not allow us to correlate the obtained data in this study to previous investigations. Antimony was not examined in mine waters in Bulgaria so far, because of its insignificant quantity in the processed ores. Taking into account its presence in the mined ores from Chiprovtsi mine and its behavior and toxicity to the environment, we decided to examine its concentration in mine waters and explore whether it could be, or is a potential pollutant of the waters in the study area. The determined total antimony concentration in 5 from 15 mine waters samples does not exceed $25 \mu\text{g/l}$, which is slightly above the maximum permissible values (MOEW 2001) and WHO guideline values (WHO 2004) (Tables 1, 2) for surface water. Actually, we did not find any data about antimony concentrations in mine, surface and drinking waters in Bulgaria at all. According to Alderton *et al.* (2005) the highest dissolved Sb concentrations occur in alkaline waters (pH 7-9) in contrast with As and are related not only to the mineralization type, but to pH. Antimony forms oxyanions and the most common is $\text{Sb}(\text{OH})_6^-$ and the solubility of this species is higher in neutral to alkaline conditions (Longmuir 1997). Similarly to other heavy metals and arsenic, antimony can be adsorbed by Fe oxides and hydroxides, and thereby disappears from the water. The fact that antimony was not determined in surface waters is an evidence that it was adsorbed very close to the adits, probably in soil, or in bed sediments inside

the adits. The determined Sb concentrations are considerably lower than those measured in Macedonia (Krstov Dol mine) (Alderton *et al.* 2005), which reached 100.3 $\mu\text{g/l}$. However, the results of August 2006 sampling compared to the September 2005 reported elsewhere (Dimitrova *et al.* 2006) show substantial difference between both dry seasons, caused probably by the higher humidity and average annual rainfall estimated for the 2005 year. The total heavy metal concentrations including Cu, Cd, Zn and Pb determined in 2006 year dry season waters are actually below the detection limits of the used ICP-AES method. These results have positive correlation with those derived in 2005 although they are much lower, as the 2005 dry season was unusually wet. The increased ionic concentration represented by the higher values of TDS and EC probably result in much higher level of precipitation processes during dry season, which in turn can explain the lack of heavy metals in the waters. They were probably included in the precipitated iron hydroxides.

Table 1 - Seasonal physico-chemical parameters and estimated heavy metal and metalloid concentrations in mine waters

N ^o	Sampling place (mine section, level)	pH	TDS (ppm)	T (°C)	EC (μScm^{-1})	Cu ($\mu\text{g/l}$)	Zn ($\mu\text{g/l}$)	Cd ($\mu\text{g/l}$)	Pb ($\mu\text{g/l}$)	As ($\mu\text{g/l}$)	Sb ($\mu\text{g/l}$)
1.	Perchinki, 940, adit	7.88	140	9.0	289	14	8	9	275	119	-
		7.53	210	8.6	445	-	-	-	-	149	-
2.	Perchinki, 900, adit	7.59	142	6.6	303	11	9	9	265	71	-
		7.73	226	7.5	451	-	-	-	-	77	-
3.	Mali Dol, 810, adit	7.40	154	8.4	315	156	332	-	-	114	4
		7.80	210	9.8	420	-	-	-	-	147	-
4.	Mali Dol, 850, adit	7.66	160	8.4	315	68	12	4	34	140	-
		7.83	242	11.4	481	24	19	-	-	144	-
5.	Zhazhkov Dol, 760, adit	8.17	225	7.2	450	-	-	-	-	-	-
		8.00	287	9.0	567	-	-	-	-	170	-
6.	Zhazhkov Dol, ore-chute	8.06	230	10.0	480	36	13	31	484	121	-
		8.18	288	11.2	562	-	-	-	-	168	-
7.	Zhazhkov Dol, 740, drainage water	7.69	-	16.1	200	-	1	5	66	12	-
		-	-	-	-	-	-	-	-	-	-
9.	Sinya Voda, 620, adit	8.54	238	8.9	649	27	15	8	171	31	-
		8.38	388	7.5	780	-	-	-	-	23	-
11.	Gnili Dol, 570, lake in front the adit	8.29	148	14.7	300	17	7	10	277	9	17
		8.49	175	17.5	350	-	-	-	-	-	-
12.	Velin Dol, 545, adit	7.90	202	11.6	405	14	8	11	264	7	-
		7.94	202	12.6	416	-	-	-	-	-	-
13.	Lukina Padina, 495, adit	8.47	275	9.4	541	19	25	47	1456	34	18
		8.26	343	12.3	680	-	-	-	-	20	-
14.	Dolich, 415, adit	8.05	240	11.5	492	11	10	11	431	8	25
		8.06	305	13.0	605	-	-	-	-	13	-
15.	Dolich, 395, drainage water	8.30	237	14.0	474	8	7	8	259	7	21
		8.76	265	27.1	534	-	-	-	-	-	-

Values for each sampling point on first line – May 2006, second line – August 2006; - Element not found at detection limits: Cu – 5 $\mu\text{g/l}$, Zn – 2 $\mu\text{g/l}$, Pb and Sb – 10 $\mu\text{g/l}$, As – 7 $\mu\text{g/l}$;

Abbreviations: pH – measure of the acidity and alkalinity, TDS – total dissolved solids, T – temperature, EC - conductivity

Lead concentration variations during May sampling vary between 34 $\mu\text{g/l}$ (Mali Dol) and 1456 $\mu\text{g/l}$ (Lukina Padina) to absolute absence in all samples during August, which most probably is controlled by the entire water ionic composition during the wet and dry seasons. Of particular significance are the SO_4^{2-} and HCO_3^- concentrations, which in principle control the lead content in mine waters (Blowes and Ptacek 1994). Most likely, lead was included in the bed sediments inside the adits.

Table 2 - WHO (World Health Organization) guideline values and MOEW (Ministry of Environment and Water) maximum permissible values (MPV) of Cu, Cd, Zn, Pb, As and Sb for drinking and surface water

Organization, guideline values	pH	T (°C)	EC (μScm^{-1})	Cu ($\mu\text{g/l}$)	Zn ($\mu\text{g/l}$)	Cd ($\mu\text{g/l}$)	Pb ($\mu\text{g/l}$)	As ($\mu\text{g/l}$)	Sb ($\mu\text{g/l}$)
MOEW (1986), MPV, surface water, 2 nd category	6.0-8.5	20.0	1300	100	5000	100	50	50	-
MOEW (2000), MPV, waste water	6.0-9.0	-	-	500	2000	100	200	100	-
MOEW (2001), MPV, drinking water	6.5-9.5	20.0	2000	2000	5000	5	10	10	5
WHO (2004), guideline values, drinking water	<8.0	-	-	2000	3000	3	10	10	20

On the contrary, arsenic concentrations remain high and almost constant during the monitored seasons including September 2005 (Dimitrova *et al.* 2006). Some authors consider arsenopyrite as relatively stable mineral under oxidizing conditions, near neutral pH and comparatively high Eh (Jambor 1994). However, there are numerous papers concerning oxidation processes of arsenopyrite and pyrite. Through oxidation in fresh air, water, or even mine waters, coatings of Fe (III) oxyhydroxides up to 50 Å thick are formed on the minerals surfaces (Nesbitt and Muir, 1998). These coatings commonly include hematite, goethite, and an overlayer composed also by arsenate ($\text{AsO}[\text{OH}]_3$), or FeAsO_4 , and even the reduced arsenic species ($\text{As}[\text{OH}]_3$), or FeAsO_3 . These coatings did not offer protection of continuous leaching of arsenic from the bulk. Thus, arsenopyrite still dissolves in water through oxidation and leaching processes and its concentration reaches up to 200-300 $\mu\text{g/l}$. Arsenopyrite decomposes following the simplified reaction: $\text{FeAsS} + 7\text{H}_2\text{O} = \text{Fe}^{2+}(\text{aq}) + \text{H}_3\text{AsO}_3(\text{aq}) + 11\text{H}^+ + 11\text{e}^- + \text{SO}_4^{2-}$. As far as dissolution is a function of the temperature, substance chemistry, pH and Eh, could be assumed, that at 8-10 °C and pH – 7-8 of the water in the underground workings, the arsenopyrite dissolution would be lower than that calculated for 25 °C by Craw *et al.* (2003). According to this study, the dissolution is probably enough to produce about 100-170 $\mu\text{g/l}$ dissolved $\text{As}^{3+}(\text{aq}) + \text{As}^{5+}(\text{aq})$ in the waters, as it was determined in mine waters from Perchinki, Mali Dol and Zhazhkov Dol mine sections.

5.2. Seasonal concentration variations of heavy metals and metalloids in stream, river and drinking waters

Chiprovtska Ogosta river has a relatively small and intermittent flow. The average measured quantities at Chiprovtsi station are as follows: maximum values during April - May - 1.9 m^3/s and minimum during August - 0.2 m^3/s .

The concentrations of the dissolved heavy metals and metalloids decrease during the wet season reaching values below the detection limits of the ICP-AES method, as being diluted by the water. This is not valid for the arsenic concentrations, which vary between <10 and 297 $\mu\text{g/l}$. The highest arsenic concentration (297 $\mu\text{g/l}$) is determined at the clarified waters collector outlet of the Golyam Bukovets tailing impoundment and it appears to be three times higher than the permissible value for hazardous and harmful substances (100 $\mu\text{g/l}$) in effluent waste waters discharged into the river systems (Table 2, MOEW 2000). The arsenic values determined in the natural waters collector outlet of the Mechi Dol tailing impoundment nearly reach the permissible limit. The Chiprovtska Ogosta river basin is signed to a second water stream category in Bulgaria (MOEW 1986), i.e.

the surface waters are not applicable for drinking water-supply, but are suitable for irrigation, fish farming, watering and for industrial purposes. As it is shown in Table 3 and Figure 2, the arsenic concentrations estimated for all sampling points along the river downstream during wet season exceed the guideline value for 2nd surface waters category nearly 2.7 times. Especially noticeable is the concentration increase, found out in the river waters after entering the town of Chiprovtsi, which is probably due to the flow from the left tributaries Rupski Dol and Gnili Dol, which pass through the mining areas. It apparently decreases at the Montana artificial lake near Gorno Tserovene village, reaching levels below the guideline values for the second surface waters category (44 µg/l) (Fig. 2).

The self-purification processes are considerably reduced through the dry season (August 2006) and arsenic concentrations determined in all sampling points exceed the guideline values from 1.1 to 3.46 times. The highest concentration estimated in the river waters is at the point near Zhelezna village.

The three tailing impoundments in the mines vicinity significantly contribute in river waters pollution, particularly regarding arsenic concentration. Among them, the natural waters collector of Mechi Dol tailing impoundment releases the highest contents of dissolved arsenic - 621 µm/l, which 6 times exceeds the guideline values for waste waters. Nevertheless, arsenic rather rapidly precipitates or is adsorbed by the channel sediments not very far from the outlet – at about 2 km, as the concentration decreases to 85 µm/l before flowing into Martinovo artificial lake, as it is shown on Figure 3. On the other hand, Chiprovtsi tailing impoundment directly contributes in river pollution by metal remobilization processes caused by its immediate riverside location. In addition, Sinya Voda stream, which rinses the mine waste material near the adit, increases the arsenic concentration in Chiprovaska Ogosta river up to 3 times above the guideline values. Most of the determined arsenic concentrations in surface waters result from oxidation and remobilization processes occur in the tailing impoundments, which contain significant amounts of arsenopyrite, pyrite and other sulphides (Vesselinov *et al.* 1996).

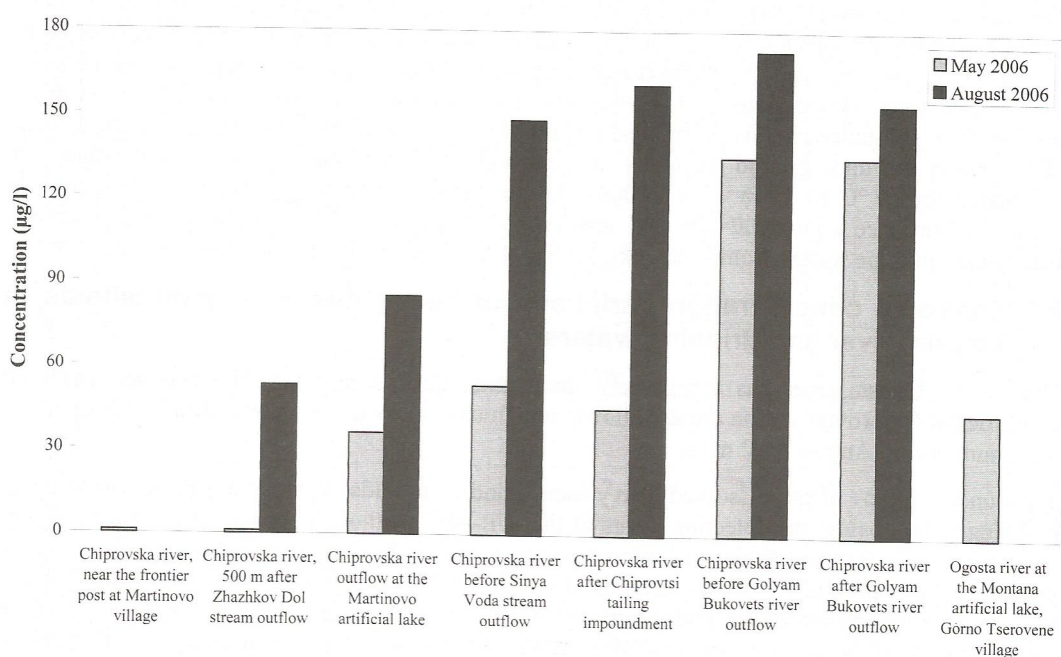


Figure 2 - Total arsenic concentrations determined at certain sampling points along the Chiprovaska Ogosta downstream during May and August 2006

Table 3 - Physico-chemical parameters and total arsenic concentration in surface waters during May and August 2006 sampling

№	Sampling place	pH		TDS (ppm)		T (°C)		EC (µScm ⁻¹)		As (µg/l)	
		May	August	May	August	May	August	May	August	May	August
8.	Yavorov Dol, stream	8.15	7.71	94	126	11.3	17.6	188	253	-	10
10.	Rupski Dol, stream	8.15	8.25	150	151	15.8	17.9	294	306	14	-
18.	Chiprovska river, near the frontier post at Martinovo village	7.01	7.07	18	33	13.1	14.6	38	65	<10	-
19.	East tributary flowing into Mechi Dol tailing impoundment	7.73	7.56	42	62	14.1	13.1	83	124	<10	<10
20.	West tributary flowing into Mechi Dol tailing impoundment	7.73	7.53	42	59	15.3	14.7	84	119	<10	<10
21.	Natural waters collector outlet, Mechi Dol tailing impoundment	7.25	7.22	145	248	13.1	12.9	300	503	109	621
22.	Chiprovska river, 500 m after Zhazhkov Dol stream outflow	7.80	7.92	30	83	11.5	15.3	65	181	<10	53
23.	Chiprovska river outflow at the Martinovo artificial lake	7.97	7.85	42	96	11.7	17.8	85	196	36	85
24.	Chiprovska river before Sinya Voda stream outflow	7.92	7.93	43	94	16.0	18.5	86	188	53	148
25.	Sinya Voda stream mouth	7.99	8.02	146	224	15.0	17.3	290	450	61	68
26.	Chiprovska river after Chiprovtsi tailing impoundment	7.80	8.02	57	102	15.5	18.1	116	202	45	161
27.	Natural waters collector inlet of Golyam Bukovets tailing impoundment	7.65	7.54	78	105	17.5	19.7	155	211	<10	<10
28.	Lake in Golyam Bukovets tailing impoundment	8.23	9.19	101	105	21.2	25.2	206	211	148	351
29.	Natural waters collector outlet of Golyam Bukovets tailing impoundment	7.94	7.91	210	267	14.6	16.5	420	533	297	183
30.	Clarified waters collector outlet of Golyam Bukovets tailing impoundment	7.46	7.45	387	501	11.7	12.9	774	1001	101	228
31.	Chiprovska river after Golyam Bukovets river outflow	8.16	8.16	172	187	15.5	20.2	344	376	135	154
32.	Golyam Bukovets river mouth	8.09	7.96	198	250	14.0	18.0	198	520	57	85
33.	Chiprovska river before Golyam Bukovets river outflow	7.82	7.90	136	126	15.6	18.7	272	253	135	173

- The total concentrations of Cu, Cd, Zn, Pb and Sb are not included in the table, because they were not detected at detection limits for analyses 8 and 10: Cu – 5 µg/l, Zn – 2 µg/l, Pb and Sb – 10 µg/l, As – 7 µg/l; for analyses 18-33: Cu, Cd and Zn – 1 µg/l, Pb and Sb – 10 µg/l, As – 6-14 µg/l

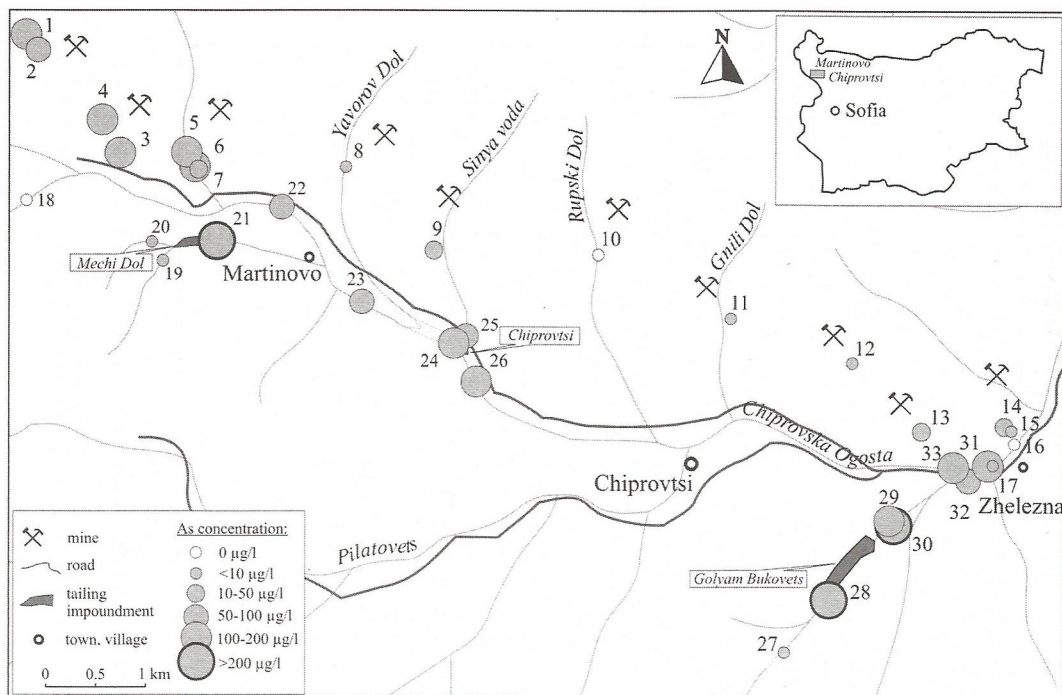


Figure 3 - A view of the studied area showing arsenic levels ($\mu\text{g/l}$) in mine, surface and drinking water

In comparison to previous studies carried out during August 2005 (Cholakova *et al.* 2006), an obvious trend of arsenic concentration elevation is outlined in Chiprovtska Ogosta river on leaving the town of Chiprovtsi. However, they cannot be linked to concrete impact of the tailing impoundments and mining workings. Thus, a detailed sampling along the river downstream, and all left tributaries between Chiprovtsi tailing impoundment and Golyam Bukovets river mouth has to be done, including total ionic composition monitoring as well.

The investigation of the drinking water samples did not show any concentrations of heavy metals and arsenic (Table 4), i.e. drinking water from the public fountains in Zhelezna village is absolutely harmless to humans. Further monitoring is still needed assuring our statement.

Table 4 - Physico-chemical parameters and total Cu and Zn concentration in drinking waters during May 2006 sampling

№	Sampling place	pH		TDS (ppm)		T ($^{\circ}\text{C}$)		EC (μScm^{-1})		Cu ($\mu\text{g/l}$)	Zn ($\mu\text{g/l}$)
		May	August	May	August	May	August	May	August	May*	May*
16.	Zhelezna village, near Dolich mine section, drinking water	-	7.79	-	283	-	16.8	-	578	-	-
17.	Zhelezna village, drinking water	7.20	7.06	105	140	10.6	17.6	212	284	4	6

* - As, Sb, Pb, Cd were not detected during the monitoring, while Cu and Zn were detected only during May 2006 sampling

6. Conclusions

The measured physico-chemical parameters of mine and surface waters during dry season also show high conductivity and TDS values. The pH values in all sampled waters remain near neutral during the monitored period showing slight elevation during dry season. This fact, resulting from the sufficient amount of CO_3^{2-} , HCO_3^- , OH^- in the water, which come from the host rocks (siderite marbles), confirms that acid mine drainage from the adits and tailing impoundments could not be expected and there is no need of further activities to limit such effect.

The most hazardous element in mine and surface waters is arsenic. Its concentrations in mine waters show insignificant variability between wet and dry seasons, but remain almost equal. The highest As concentrations (up to $170 \mu\text{g/l}$) in mine waters are observed in the western part of the mining area (Perchinki, Mali Dol and Zhazhkov Dol mine sections). In surface waters arsenic concentration is ranging between 60 and $80 \mu\text{g/l}$, but abruptly increases above $170 \mu\text{g/l}$ instantly after each tailing impoundment water collector outflow to the river. The highest arsenic concentrations (from 228 to $621 \mu\text{g/l}$) are observed during dry season, resulting from the low flow in the whole river system. The determined high arsenic concentrations in river waters (several times exceeding the guideline values) categorize them as unsuitable for irrigation.

The other heavy metals and antimony are not found in surface waters in concentrations above the guideline values and do not pose a threat to the waters. In mine waters, these metals occur in much higher values, especially lead, which concentrations vary between $1456 \mu\text{g/l}$ and $<10 \mu\text{g/l}$ during wet and dry seasons, respectively. As lead concentration in water very much depends on the anionic composition, particularly SO_4^{2-} and HCO_3^- , during dry season it is easily adsorbed by the channel sediments and soil, or precipitates as sulphate (anglesite). Thus, it is considered as a serious contaminant to soil and vegetation.

This study did not record any contamination in drinking water, even with arsenic, thus using drinking water from the public fountains in Zhelezna village do not pose a threat to the population.

The concentrations of As, Cu, Cd and Pb are significantly decreased in comparison to these measured during mining (Kotsev 2001, 2003), but still remain very high regarding the guideline values. This means, that sulphide oxidation and metal remobilization processes are still running in the underground workings and tailing impoundments and they remain a constant source of heavy metal and metalloid contamination mainly for the water bodies and riverside agricultural land. Thus we consider, that the remediation activities carried out in the mining area are not very successful and a seasonal monitoring of the pollutants concentration for several years will show in what extent a pollution of waters can be expected in the area and what preventive measures have to be taken to restrict such contaminations.

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8. References

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